Contents lists available at ScienceDirect





Journal of Hazardous Materials

journal homepage: www.elsevier.com/locate/jhazmat

Gamma radiolysis of alachlor aqueous solutions in the presence of hydrogen peroxide

Dongkyu Choi^a, O-Mi Lee^a, Seungho Yu^a, Seung-Woo Jeong^{b,*}

^a Advanced Radiation Technology Institute, Korea Atomic Energy Research Institute, Jeongeup 580-185, Republic of Korea ^b Department of Environmental Engineering, Kunsan National University, Kunsan 573-701, Republic of Korea

ARTICLE INFO

Article history: Received 15 July 2010 Received in revised form 4 August 2010 Accepted 7 August 2010 Available online 14 August 2010

Keywords: Alachlor Gamma irradiation Hydrogen peroxide (H₂O₂) Biodegradability

ABSTRACT

The enhanced effect of gamma irradiation with hydrogen peroxide (H_2O_2) for alachlor degradation in an aqueous solution was first investigated in this study. The combination of gamma irradiation and H_2O_2 led to an enhanced effect, which remarkably increased the degradation efficiency of alachlor and the total organic carbon (TOC) removal. At a dose of 200 Gy, the degradation degree of the alachlor solution reached 81.7 and 99.2% under H_2O_2 concentrations of 0 and 0.1 μ M, respectively. In addition, the TOC removal efficiencies of the alachlor under initial H_2O_2 concentrations of 0, 0.5 and 1.0 μ M were 59.5, 74.8 and 83.8%, respectively, at an absorbed dose of 20 kGy. However, for higher H_2O_2 concentrations (greater than 1 μ M), the alachlor degradation was reduced because •OH radicals were scavenged by the H_2O_2 . The biodegradability of alachlor solutions prior to and after treatment by gamma irradiation was also with increasing absorbed doses.

© 2010 Elsevier B.V. All rights reserved.

1. Introduction

The increasing production and extensive use of chemical herbicides in agriculture has become an important issue due to their persistent recurrence in various sectors of our environment. The herbicide alachlor (2-chloro-2',6'-diethyl-Nmethoxymethylacetanilide) is widely used to control broadleaf and grass weeds in agricultural crops [1,2]. It has been detected from 0.1 to 10 μ g/L in groundwater and surface water in the United States, Europe, Japan, and China [3,4]. Due to its toxicity, recalcitrance, and potential adverse health effects, alachlor has been classified as a B2 carcinogen by the U.S. Environmental Protection Agency (USEPA) and is one of the priority substances (PS) listed by the European Commission (EC) [2,3,5]. Alachlor, a extremely toxic endocrine disruptor, has been restricted to a maximum of 2 μ g/L in drinking water by the USEPA [3,4,6].

Alachlor is a persistent organic pollutant with high toxicity to many organisms due to its strong aromaticity and high chemical stability. Conventional biological and physico-chemical processes are inadequate for its degradation in contaminated water [2,7]. Therefore, efficient and cost effective treatment methods to remove refractory organic pollutants have been suggested. In recent years, advanced oxidation processes (AOPs) have received significant attention because they are based on the generation and use of highly reactive hydroxyl radicals (•OH) for the removal of various organic pollutants from water and wastewater [8,9]. Various AOPs, such as ozonation [3], TiO₂ and UV process [10,11], photo-Fenton oxidation [4,5], ferrate oxidation [12], and ultrasonic treatment [13], have been proposed for the removal of alachlor. These processes still have limitations such as low removal efficiency and high operational cost. Gamma irradiation is considered to be an effective process for removal of refractory organic pollutants.

Gamma irradiation has proved to be a promising solution for the degradation of refractory substances such as fluorine [14], methyl tert-butyl ether (MTBE) [15], nitrobenzene [16] and cefaclor [17]. As a consequence of the interaction between the gamma irradiation and the aqueous solution, the primary species ($^{\circ}OH$, e_{aq}^{-} , and $^{\circ}H$) and molecular products (H₂, H₂O₂) are generated according to the following equation [18]:

$$\begin{split} H_2 O &\rightarrow 2.7 e_{aq}^- + 0.6 H^\bullet + 2.8^\bullet OH + 0.45 H_2 + 0.7 H_2 O_2 \\ &+ 3.2 H_{aq}^+ + 0.5 OH_{aq}^- \end{split} \tag{1}$$

The numbers in Eq. (1) are the *G*-values (#species/100 eV) of each species. These reactive species and radicals react with the target compounds and decompose of the solutes present in wastewater. The chemical reactions of the three reactive species ($^{\circ}$ OH, e_{aq}^{-} , and $^{\circ}$ H), particularly the $^{\circ}$ OH radicals, are highly reactive and are key importance with regard to gamma radiation and electron beam processes in aqueous solutions. Recently, Da Silva and Vieira [6] reported on their study of radiolytic degradation

^{*} Corresponding author. Tel.: +82 63 469 4767; fax: +82 63 4694964. *E-mail address*: swjeong@kunsan.ac.kr (S.-W. Jeong).

^{0304-3894/\$ –} see front matter s 2010 Elsevier B.V. All rights reserved. doi:10.1016/j.jhazmat.2010.08.037

of alachlor in aqueous solution by gamma radiation. The result showed that gamma radiation was effective for alachlor degradation in an aqueous solution. However, a high dose of 5.0 kGy was required for the degradation of alachlor. This called for an innovative means of reducing the absorbed dose for practical use. Thus, we employ the combination process of gamma irradiation and H_2O_2 to enhance alachlor degradation.

This study focused on the combination of gamma irradiation and H_2O_2 as an effective treatment method for the degrading of alachlor found in the environment. The main objectives were (1) to study the degradation kinetics of alachlor; (2) to investigate the enhanced effect of gamma irradiation and H_2O_2 on the radiolytic decomposition and mineralization of alachlor in aqueous solutions; (3) to evaluate the biodegradability of the irradiated solutions after gamma irradiation.

2. Materials and methods

2.1. Chemicals

Alachlor (99.2%, HPLC grade), the structure of which is shown in Fig. 1, was obtained from Sulpelco Inc. (Bellefonte, PA, USA). Hydrogen peroxide (H_2O_2 , Showa Chemical Co., Ltd., Japan) was used as a 30% (w/w) solution. Acetonitrile used for HPLC analysis was of HPLC grade and purchased from J.T. Baker (USA). All the solutions were prepared using Milli-Q purified water (Millipore Inc., USA).

2.2. Irradiation sources

Gamma irradiation was performed with a high-activity ⁶⁰Co source (Nordion Inc., Canada) at the Korea Atomic Energy Research Institute (Jeongeup, Korea). The radioactivity of the source was around 9.85×10^{15} Bq (= 266,200 Ci). Aqueous alachlor solutions were irradiated in 15 mL screw-cap-bottles without any headspace. The solutions were prepared 24 h prior to irradiation, sealed with screw caps to avoid contact with air and stored at 4 °C. All the samples were allowed to reach equilibrium at atmospheric pressure and room temperature (20 ± 2 °C) before being irradiated.

2.3. Analytical methods

The concentration of alachlor in the aqueous solution was measured using high performance liquid chromatography, using an Agilent 1200 Series HPLC (Agilent Technologies, Santa Clara, CA, USA), equipped with a UV absorbance detector and a Discovery C18, 5 μ m, 150 mm × 4 mm column (Supelco, USA). The eluent consisted of acetonitrile/distilled water (60/40, v/v) and the flow rate was 1.0 mL/min. The wavelength for alachlor was set at 230 nm and



Fig. 1. Chemical structure of alachlor.

the column temperature was kept at 30 °C. The total organic carbon (TOC) in the aqueous samples was determined using a Shimadzu TOC-VCPH analyzer (Shimadzu Co., Japan).

2.4. Biodegradability in the Closed Bottle Test (OECD 301 D)

The CBT (Closed Bottled Test) is recommended as a first, simple test for the assessment of the biodegradability of organic compounds [19]. It was performed with distilled water at room temperature $(20 \pm 2 \circ C)$, according to test guidelines [19]. The inoculum was taken from the effluent of a municipal wastewater treatment plant (Jeongeup, Korea). During the CBT, the incubated alachlor samples were monitored by measuring the oxygen concentration (OxiTop, WTW Weilheim, Germany).

3. Results and discussion

3.1. Degradation kinetics of alachlor

Radiation-induced degradation of organic pollutants is initiated by primary species or radicals such as •OH radicals, e_{aq}^- and •H, produced from water radiolysis (Eq. (1)). H_2O_2 and H_2 do not take part in the degradation reactions because of their low reactivity and yields. In the presence of O_2 , the reducing species, •H and e_{aq}^- , are converted into oxidizing species, HO_2^{\bullet} and $O_2^{\bullet-}$, respectively (Eqs. (2)–(4)) [18,20,21].

$$H^{\bullet} + O_2 \rightarrow HO_2^{\bullet} \quad k = 2.1 \times 10^{10} \,\mathrm{M}^{-1} \,\mathrm{s}^{-1}$$
 (2)

$$e_{aq}^{-} + O_2 \rightarrow O_2^{\bullet -} \quad k = 1.9 \times 10^{10} \,\mathrm{M}^{-1} \,\mathrm{s}^{-1}$$
 (3)

$$HO_2^{\bullet} \leftrightarrow H^+ + O_2^{\bullet -} \quad pK = 4.7 - 4.8 \tag{4}$$

The concentration of •OH radicals is higher than that of HO_2^{\bullet} and $O_2^{\bullet-}$. Thus, the •OH radicals present in aerated solutions have been shown to contribute significantly to the radiolytic degradation of alachlor.

In this study, alachlor aqueous solutions of 11.5, 18.9, 29.0 and $45.4 \,\mu\text{M}$ were irradiated at 0–500 Gy. The results showed that the concentration of alachlor decreased exponentially with the increase of the absorbed doses, indicating that alachlor degradation followed a pseudo-first-order kinetic model, which can be described by the following equations [22,23]:

$$C = C_0 \exp(-dD) \tag{5}$$

$$-\ln\left(\frac{C}{C_0}\right) = dD,\tag{6}$$

where C_0 is the initial concentration of the alachlor before irradiation; *C* is the concentration after irradiation; *d* is the dose constant in units of a reciprocal dose; and *D* is the absorbed dose.

Fig. 2 shows the degradation of alachlor at different initial concentrations with increasing absorbed doses. Alachlor decreased with increasing absorbed dose. For a dose of 150 Gy, the removal efficiencies for initial aqueous alachlor concentrations of 11.5, 18.9, 29.0 and 45.4 μ M were 97.7, 91.2, 79.2 and 69.3%, respectively. Fig. 3 and Table 1 show the radiolytic degradation rate constants of alachlor at different initial concentrations, indicating a dependence on the initial alachlor concentrations of alachlor. The relationship

Table 1
Change in dose constant (d) with increasing the initial concentration of alachlor.

[alachlor] ₀ (µM)	Dose constant, d (Gy ⁻¹)	R^2
11.5	0.0242	0.9915
18.9	0.0173	0.9855
29.0	0.0108	0.9980
45.4	0.0084	0.9951



Fig. 2. Radiolytic degradation of alachlor at different initial concentrations with increasing absorbed doses.

between the dose constant and the initial concentrations of the pollutants was also observed in the study of gamma radiolysis for pentachlorophenol by Xue and Wang [23] and of PCB by Mincher et al. [24]. Therefore, the absorbed dose required to remove the alachlor concentration in aqueous solution under the same experimental conditions can be easily calculated.

3.2. The effect of added H_2O_2 on alachlor degradation

 H_2O_2 is a well-known source of •OH radicals, which is widely used to remove refractory organic pollutants in wastewater [25,26]. The effect of H_2O_2 on alachlor (40 µM) degradation was investigated for the different absorbed doses. Fig. 4 shows the concentration of residual alachlor as the efficacy of the concentration of H_2O_2 on the absorbed doses. The concentrations of added H_2O_2 were varied from 0 to 10 µM. As shown in Fig. 4, the residual alachlor concentration had decreased significantly as the concentration of H_2O_2 increased up to 1.0 µM, but as more H_2O_2 was added, the residual alachlor concentration gradually increased. H_2O_2 can react with the e_{aq}^- and •H produced by the water radiolysis (Eq. (1)), converting them into more •OH radicals, as shown in Eqs. (7) and (8) [26,27].

$$e_{ag}^{-} + H_2O_2 \rightarrow {}^{\bullet}OH + OH^{-} \quad k = 1.1 \times 10^{10} \,\mathrm{L}\,\mathrm{mol}^{-1}\,\mathrm{s}^{-1}$$
 (7)

•H + H₂O₂
$$\rightarrow$$
 •OH + H₂O $k = 9.0 \times 10^7 \,\mathrm{L \, mol^{-1} \, s^{-1}}$ (8)

Thus, the combination of H_2O_2 with gamma radiation could significantly enhance the degradation of alachlor. However, further increase of H_2O_2 (>1.0 μ M) could compete with organic compounds for •OH radicals (Eqs. (9) and (10)) and therefore increase the resid-



Fig. 3. Initial concentration dependency of the radiolytic degradation rate constant (dose constant) by gamma irradiation.



Fig. 4. Residual alachlor after gamma irradiation with H_2O_2 ([alachlor]_0 = 40 $\mu M,$ $[H_2O_2]_0$ = 0–10 $\mu M).$

ual concentration of alachlor (Fig. 4) [26,28].

•OH + H₂O₂
$$\rightarrow$$
 HO₂• + H₂O $k = 2.7 \times 10^8 \,\mathrm{L\,mol^{-1}\,s^{-1}}$ (9)

$${}^{\bullet}\text{OH} + {}^{\bullet}\text{OH} \to \text{H}_2\text{O}_2 \quad k = 1.1 \times 10^{10} \,\text{L}\,\text{mol}^{-1}\,\text{s}^{-1} \tag{10}$$

The relationship between gamma radiation and H_2O_2 on the degradation effect of the target organic pollutants (*p*-nitrophenol, dyes (red KE-3B and blue XBR)) has been reported previously by Yu et al. [29] and Wang et al. [27]. The radiolytic degradation efficiencies of the pollutants increased with the increasing initial H_2O_2 concentration, consistent with our experimental results. However, an optimal ratio of H_2O_2 concentration for alachlor degradation was observed in our study. When H_2O_2 concentration exceeds the optimal ratio, it could play an important role as a •OH radical scavenger as shown in Eqs. (9) and (10).

3.3. TOC removal of alachlor

In order to investigate the effects of H₂O₂ on the mineralization of alachlor by gamma irradiation, the TOC values of the irradiated solutions were analyzed. Fig. 5 shows the TOC removal with increasing absorbed doses at different initial H2O2 concentrations. When the initial H₂O₂ concentrations were in the range of $0-1.0 \,\mu$ M, more TOC was removed as the concentration of H₂O₂ increased. For an absorbed dose of 20 kGy, the TOC removal rates from the alachlor solutions with initial H₂O₂ concentrations of 0, 0.5 and 1.0 µM were approximately 59.5, 74.8 and 83.8%, respectively. When irradiated with the same absorbed doses, a higher H₂O₂ concentration in the solution showed more effective TOC removal efficiency. Similar results were reported for the TOC removal of chlorophenols [25] and *p*-nitrophenol [29]. As shown in Eqs. (7) and (8), this could be due to H_2O_2 promoting the mineralization of alachlor. However, the TOC in the irradiated alachlor solutions with different initial H₂O₂ concentrations was not completely removed even for an absorbed dose of 20 kGy (Fig. 5). Li et al. [3] and Qiang et al. [30] reported that organic acids such as formic, acetic, oxalic and propionic acid were produced as organic by-products after degradation of alachlor by O_3/H_2O_2 and O_3 . It is suggested that the reduction of the TOC removal efficiency could be due to the



Fig. 5. TOC removal efficiencies of alachlor (40 μ M) with increasing absorbed doses at different initial H₂O₂ concentrations.

remaining organic by-products produced from the alachlor decomposition. These by-products were not completely mineralized.

3.4. Biodegradability of alachlor

The biodegradability of alachlor solutions $(40 \,\mu\text{M})$ prior to and after treatment by gamma irradiation was assessed according to the CBT (Closed Bottled Test). Aqueous solutions of 40 μ M alachlor were irradiated with doses of 0, 200, 400 and 1000 Gy, and the biodegradability results are shown in Fig. 6. Note that 'Control' of Fig. 6 says 40 μ M alachlor solutions without inoculums. As shown in Fig. 6, the tested alachlor was readily biodegraded in the CBT and the biodegradability of the irradiated solutions were enhanced as the absorbed doses were increased. The irradiated alachlor solutions had been biodegraded by 63.1–72.0% during the test period of 28 days. Thus, they can all be classified as readily biodegradable. Torres



Fig. 6. Biodegradability of alachlor aqueous solutions (40 μ M) irradiated by gamma irradiation; 'Control': 40 μ M alachlor solutions without inoculum.

et al. [13] also reported that the biodegradability test (BOD5/COD) results showed a significant enhancement in the biodegradability of the alachlor solution after ultrasonic treatment. Therefore, our experimental results indicate that the gamma irradiation process transforms alachlor into more biodegradable and less toxic by-product.

4. Conclusions

Based on the experiments on the radiolytic degradation of alachlor and on the effect of H_2O_2 , we can conclude this study with the following points:

- (1) Alachlor was effectively degraded by gamma irradiation. Its degradation followed pseudo-first-order kinetic model, and the dose constant *d* increased with decreasing initial concentrations of alachlor.
- (2) An enhanced effect was found on the degradation of alachlor by gamma irradiation with H₂O₂. However, increasing the initial concentration of H₂O₂ in the solution reduced the alachlor degradation efficiency because H₂O₂ acted as a scavenger of the •OH radicals at high H₂O₂ concentrations.
- (3) In the presence of H_2O_2 , the removal efficiency of TOC was enhanced significantly as the absorbed dose was increased. At an absorbed dose of 20 kGy, TOC removal efficiency of the alachlor with initial H_2O_2 concentration of 1.0 μ M was approximately 83.8%.
- (4) The biodegradability of irradiated alachlor samples was enhanced with increasing absorbed doses, indicating that a significant portion of the non-biodegradable alachlor is transformed into biodegradable by-products by gamma irradiation.

Acknowledgement

This research was supported by the Nuclear R&D program of the Ministry of Education, Science and Technology of Korea.

References

- J. Qu, H. Li, H. Liu, H. He, Ozonation of alachlor catalyzed by Cu/Al₂O₃ in water, Catal. Today 90 (2004) 291–296.
- [2] X. Wang, Y. Zhang, Degradation of alachlor in aqueous solution by using hydrodynamic cavitation, J. Hazard. Mater. 161 (2009) 202–207.
- [3] H.-Y. Li, J.-H. Qu, H.-J. Liu, Decomposition of alachlor by ozonation and its mechanism, J. Environ. Sci. 19 (2007) 769–775.
- [4] H. Katsumat, S. Kaneco, T. Suzuki, K. Ohta, Y. Yobiko, Photo-Fenton degradation of alachlor in the presence of citrate solution, J. Photochem. Photobiol. A: Chem. 180 (2006) 38–45.
- [5] M.M. Ballesteros Martin, J.A. Sanchez Perez, J.L. Garcia Sanchez, L. Montes de Oca, J.L. Casas Lopez, I. Oller, S. Malato Rodriguez, Degradation of alachlor and pyrimethanil by combined photo-Fenton and biological oxidation, J. Hazard. Mater. 155 (2008) 342–349.

- [6] M.P. Da Silva, E.M. Vieira, Degradation of alachlor herbicide by gamma radiation from cobalt-60 in aqueous and alcohol solution, J. Radioanal. Nucl. Chem. 281 (2009) 323–327.
- [7] L. Calvo, M.A. Gilarranz, J.A. Casas, A.F. Mohedano, J.J. Rodriguez, Hydrodechlorination of alachlor in water using Pd, Ni and Cu catalysts supported on activated carbon, Appl. Catal. B: Environ. 78 (2008) 259–266.
- [8] A.A. Basfar, K.A. Mohamed, A.J. Al-Abduly, A.A. Al-Shahrani, Radiolytic degradation of atrazine aqueous solution containing humic substances, Ecotoxicol. Environ. Saf. 72 (2009) 948–953.
- [9] M.I. Badawy, M.Y. Ghaly, T.A. Gad-Allah, Advanced oxidation processes for the removal of organophosphorus pesticides from wastewater, Desalination 194 (2006) 166–175.
- [10] M.-S. Kim, C.S. Ryu, B.-W. Kim, Effect of ferric ion add on photodegradation of alachlor in the presence of TiO₂ and UV radiation, Water Res. 39 (2005) 525–532.
- [11] C.C. Wong, W. Chu, The direct photolysis and photocatalytic degradation of alachlor at different TiO₂ and UV sources, Chemosphere 50 (2003) 981–987.
- [12] J.-H. Zhu, X.-L. Yan, Y. Liu, B. Zhang, Improving alachlor biodegradability by ferrate oxidation, J. Hazard. Mater. B 135 (2006) 94–99.
- [13] R.A. Torres, R. Mosteo, C. Petrier, C. Pulgarin, Experimental design approach to the optimization of ultrasonic degradation of alachlor and enhancement of treated water biodegradability, Ultrason. Sonochem. 16 (2008) 425–430.
- [14] P. Popov, N. Getoff, Decomposition of aqueous fluorine by r-rays and product analysis, Radiat. Phys. Chem. 69 (2004) 387–393.
- [15] L.-L. Hsieh, Y.-L. Lin, C.-H. Wu, Degradation of MTBE in dilute aqueous solution by gamma radiolysis, Water Res. 38 (2004) 3627–3633.
- [16] J.B. Zhang, Z. Zheng, G.J. Yang, Y.F. Zhao, Degradation of microcystin by gamma irradiation, Nucl. Instrum. Meth. Phys. Res. A 580 (2007) 687–689.
- [17] S. Yu, B. Lee, M. Lee, I.-H. Cho, S.-W. Chang, Decomposition and mineralization of cefaclor by ionizing radiation: kinetics and effects of the radical scavengers, Chemosphere 71 (2008) 2106–2112.
- [18] N. Getoff, Radiation-induced degradation of water pollutants-state of the art, Radiat. Phys. Chem. 47 (1996) 581–593.
- [19] OECD Guideline for Testing of Chemicals (301 D), Closed Bottle Test. Organisation of Economic Cooperation and Development, OECD, Paris, 1992.
- [20] R. Zona, S. Schmid, S. Solar, Detoxification of aqueous chlorophenol solutions by ionizing radiation, Water Res. 33 (1998) 1314–1319.
- [21] S. Weihua, Z. Zheng, A.-S. Rami, Z. Tao, H. Desheng, Degradation and detoxification of aqueous nitrophenol solutions by electron beam irradiation, Radiat. Phys. Chem. 65 (2002) 559–563.
- [22] B.J. Mincher, R.D. Curry, Considerations for choice of a kinetic fig. of merit in process radiation chemistry for waste treatment, Appl. Radiat. Isotopes 52 (2000) 189–193.
- [23] J. Xue, J. Wang, Radiolysis of pentachlorophenol (PCP) in aqueous solution by gamma radiation, J. Environ. Sci. 20 (2008) 1153–1157.
- [24] B.J. Mincher, R.R. Brey, R.G. Rodrigues, S. Pristupa, A. Ruhter, Increasing PCB radiolysis rates in transformer oil, Radiat. Phys. Chem. 65 (2002) 461–465.
- [25] J. Hu, J. Wang, Degradation of chlorophenols in aqueous solution by r-radiation, Radiat. Phys. Chem. 76 (2007) 1489–1492.
- [26] B.G. Kwon, E. Kim, J.H. Lee, Pentachlorophenol decomposition by electron beam process enhanced in the presence of Fe(III)-EDTA, Chemosphere 74 (2009) 1335–1339.
- [27] M. Wang, R. Yang, W. Wang, Z. Shen, S. Bian, Z. Zhu, Radiation-induced decomposition and decoloration of reactive dyes in the presence of H₂O₂, Radiat. Phys. Chem. 75 (2006) 286–291.
- [28] N.-Y. Gao, Y. Deng, D. Zhao, Ametryn degradation in the ultraviolet (UV) irradiation/hydrogen peroxide (H₂O₂) treatment, J. Hazard. Mater. 164 (2009) 640–645.
- [29] S. Yu, J. Hu, J. Wang, Gamma radiation-induced degradation of p-nitrophenol (PNP) in the presence of hydrogen peroxide (H₂O₂) in aqueous solution, J. Hazard. Mater. 177 (2010) 1061–1067.
- [30] Z. Qiang, C. Liu, B. Dong, Y. Zhang, Degradation mechanism of alachlor during direct ozonation and O₃/H₂O₂ advanced oxidant process, Chemosphere 78 (2010) 517–526.